



Effect of high-temperature on the swellable organically-modified silica (SOMS) and its application to gas-phase hydrodechlorination of trichloroethylene

Hyuntae Sohn^a, Gokhan Celik^a, Seval Gunduz^a, Sreshtha Sinha Majumdar^a,
Stacey L. Dean^b, Paul L. Edmiston^{b,c,**}, Umit S. Ozkan^{a,*}

^a William G. Lowrie Department of Chemical and Biomolecular Engineering, The Ohio State University, 151 W. Woodruff Avenue, Columbus, OH 43210, USA

^b ABS Materials, Inc., 1909 Old Mansfield Road, Wooster, OH 44691, USA

^c Department of Chemistry, The College of Wooster, 943 College Mall, Wooster, OH 44691, USA

ARTICLE INFO

Article history:

Received 21 October 2016

Received in revised form 18 January 2017

Accepted 15 February 2017

Available online 20 February 2017

Keywords:

Organically-modified silica (SOMS)

Hydrodechlorination (HDC)

Trichloroethylene

ABSTRACT

Pd catalysts supported on swellable organically-modified silica (SOMS) and high-temperature-treated swellable organically-modified silica (H-SOMS) were characterized and tested for gas-phase hydrodechlorination (HDC) of trichloroethylene (TCE) conditions. The high-temperature treatment on SOMS resulted in an increase in surface area and pore diameter as well as significant improvement of Pd dispersion on H-SOMS with smaller Pd particle sizes compared to the Pd/SOMS catalyst. Although the high-temperature treatment led to some alteration of the SOMS polysiloxane network, the hydrophobicity and organic vapor adsorption characteristics of SOMS were preserved. The reduction and oxidation characteristics of Pd on SOMS and HSOMS were investigated in situ using XANES technique. It was found that the Pd sites in the pores of SOMS was accessible to small molecules such as H₂, facilitating the reduction of PdO_x, whereas oxidation of metallic Pd was limited even at higher temperatures when O₂ was used. This effect was only observed over Pd/SOMS catalyst. For Pd/H-SOMS, because the pores were more widely open than Pd/SOMS, both reduction and oxidation of Pd were observed. Finally, the catalytic activity of Pd/H-SOMS for gas-phase HDC of TCE was significantly better than Pd/SOMS. When water was added to the reactant stream (TCE + H₂O), both Pd/SOMS and Pd/H-SOMS maintained its catalytic performances due to hydrophobic property of the supports.

© 2017 Elsevier B.V. All rights reserved.

1. Introduction

Trichloroethylene (TCE) has been widely used for metal degreasing processes in the U.S. industry [1,2]. However, due to its toxicity as a carcinogen [3,4], a proper treatment should be performed before releasing it to the environment. Recently, TCE has been reported as one of the most widely found volatile organic compound contaminants in groundwater [5,6], which is a major source for human drinking water. The toxicological profile for TCE published by U.S. Department of Health and Human Services reports that TCE was detected in 4.5–18% of the drinking water sources in United States with less than 30 ppb [7]. Currently, the MCL

(maximum contaminant level) set by the U.S. Environmental Protection Agency (EPA) is 5 ppb for TCE.

The most widely used groundwater remediation technologies include air stripping, [8,9] air sparging [10] and soil venting [11] that are pump-and-treat process. However, these methods only transfer TCE from liquid to gas phase, which then require a secondary vapor treatment system. This is usually done using adsorbents such as GAC (granular activated carbon) followed by an oxidation process or incinerator that produce carbon dioxide and toxins from incomplete combustion [8,11–13]. The high operating temperature is also not economically favorable. Thus, *in-situ* groundwater remediation methods such as bioremediation [14] and chemical oxidation [15,16] have been widely studied to eliminate TCE from groundwater. An alternative technique is catalytic hydrodechlorination (HDC) of TCE which is the focus of this study. The use of metal-based catalyst converts TCE into ethane and hydrogen chloride (HCl). The process can be implemented for both *in-situ* groundwater remediation and gas phase TCE vapor treatment systems.

* Corresponding author.

** Corresponding author at: Department of Chemistry, The College of Wooster, 943 College Mall, Wooster, OH 44691, USA.

E-mail addresses: pedmiston@wooster.edu (P.L. Edmiston), ozkan.1@osu.edu (U.S. Ozkan).

With respect to catalysts, noble metals have been tested and shown promising catalytic activity [17–19]. Ordóñez et al. examined eight commercial catalysts for HDC of TCE at reaction conditions of 300 °C and 50 bar in liquid phase [17]. It was found that over Pt, Pd and Ru based catalysts, high TCE conversions were obtained. In the literature, Pd metal has been the most studied catalyst for HDC of TCE [20–27]. For example, Pd supported on different types of catalyst supports have been investigated including Al₂O₃ [20], activated carbon (AC) [17], MCM-41 [28] and ZSM-5 zeolites [29]. Lowry and coworkers have reported promising catalytic performance of Pd/Al₂O₃ obtaining TCE conversion of 97% at ambient pressure and temperature in liquid phase [30]. Yu et al. demonstrated TCE conversions achieved over six different Pd catalysts supported on MCM-41, P₂₅, SiO₂, HAP, NiO and MgO under HDC of TCE conditions using formic acid instead of hydrogen gas [28]. The highest catalytic activity was obtained over Pd/MCM-41. In addition, bimetallic catalysts were employed and tested for HDC of TCE. Wong and coworkers reported a significantly higher catalytic performance using Pd/Au nanoparticles compared to Pd/Al₂O₃ [24,31]. Modification of the catalyst properties by adding Pd capping agents such as carboxymethyl cellulose (CMC) [22,32,33] and coating Pd metals with poly(dimethylsiloxane) (PDMS) [34–36] have been researched and shown high activities.

Non-noble metals such as zero valent iron (ZVI) have also been widely studied [37–40]. ZVI acts as a reducing agent in the presence of water, thus producing hydrogen gas for HDC reactions [41]. Lin and coworkers stated that among Pd/Fe, Pt/Fe, Ru/Fe and Au/Fe catalysts with 0.25% metal loading where 0.25% Pd/Fe exhibited the highest catalytic performance of all for aqueous phase HDC of TCE in the absence of hydrogen gas [42]. It was found that the production of hydrogen from water was considerably greater over Pd/Fe compared to others. To further decrease the cost of the catalyst, Ni/Fe has been used as an alternative to Pd/Fe [41,43]. More recently, molybdenum phosphide (MoP) was investigated for HDC of TCE reaction [44,45].

Gas phase HDC of TCE can be used to treat TCE vapors from large-scale plants where incinerator or chemical oxidation processes are normally utilized [46]. The concentration of TCE vapors especially in degreasing industry ranges from 1 to 100 ppm [7]. In the literature, the reaction is performed at relatively higher temperatures and has been the focus of fewer studies compared to those in the liquid-phase [19,47,48]. Meshesha et al. reported 80% ethylene selectivity from HDC of TCE using Pd/NiMgAl mixed oxide catalyst at 573 K in a fixed bed reactor [48]. Finocchio and coworkers⁴⁶ tested several metal-free catalysts to treat TCE vapor in humid and dry reaction conditions. They obtained complete conversion of TCE over γ -alumina at temperature of 820 K.

Previously, we have reported on a new class of materials, swellable organically-modified silica (SOMS), as catalyst support for Pd used in HDC of TCE [49]. SOMS is an extremely hydrophobic and adsorptive material, which can adsorb various types of organic compounds from air and water. When sufficient adsorption takes place, pore filling can initiate the physical expansion of the SOMS volume, called as “swelling”. The increase in volume comes with an increase in pore volume and surface area [50]. The swelling mechanism is reversible, thus, when the adsorbed molecules are liberated from SOMS, the physical volume of SOMS returns to its original size. Steady-state liquid phase reaction experiments showed a higher activity and stability for Pd/SOMS catalysts compared to commercial Pd/Al₂O₃ catalysts. The higher catalytic activity over Pd/SOMS was attributed to the swelling behavior of SOMS, which only adsorbed TCE molecules from the solution (hydrophobicity) and concentrated them in the vicinity of the Pd active sites inside the pores [49]. Additionally, the Pd active sites were protected by the strong hydrophobicity of SOMS from poisoning by sulfur anions dissolved in water [49]. It was observed

that the Li₂S poisoned-1% Pd/SOMS maintained its catalytic activity compared to fresh 1% Pd/SOMS whereas Li₂S poisoned-1% Pd/Al₂O₃ catalyst was completely deactivated. Also, Pd supported on SOMS was seen to be more resistant to leaching in acidic medium [49]. When both Pd/SOMS and Pd/Al₂O₃ were treated with HCl, most of the Pd particle were leached out from the Al₂O₃ support while Pd particles were still present on the SOMS. Kopinke et al. reported similar results using silicone polymer coated Pd/Al₂O₃ catalysts where hydrophobicity of the material facilitated better protection of the Pd active sites from ionic poisoning sources such as bisulphite for 24 h [35].

Although Pd/SOMS showed promising activity in the liquid phase, in the gas phase, the activity was much lower than that of Pd/Al₂O₃. This was attributed to the fact that in the gas phase, the SOMS matrix adsorbs less total mass of TCE than in aqueous solution where adsorption of the TCE from water into by the hydrophobic pore structure is favored. In the gas phase, the swollen state is not achieved, making the active sites inside the pores inaccessible.

In this article, we report the use of SOMS treated at high temperatures under an inert environment (H-SOMS) prior to the Pd impregnation step. This treatment was to improve Pd accessibility of SOMS for the gas-phase HDC of TCE. The evolution of SOMS during high-temperature treatment was examined using temperature programmed decomposition (TPD) technique [29]. Si Nuclear magnetic resonance (NMR) and Infrared (IR) Spectroscopy techniques were used to understand the chemical structure of H-SOMS. High Angle Annular Dark Field (HAADF) Scanning Transmission Electron Microscopy (STEM) and CO chemisorption methods were employed to obtain particle size and Pd dispersion. The change in Pd oxidation state during reduction and oxidation process were followed using *in-situ* X-ray Absorption Near Edge Structure (XANES) Spectroscopy. Lastly, gas-phase HDC of TCE was conducted at 100, 150 and 200 °C to acquire TCE and HDC conversion and product distribution. All the characterization techniques and activity testing were also performed over regular Pd/SOMS for comparison.

2. Experimental

2.1. Catalyst preparation

A detailed sol-gel synthesis method to prepare SOMS was reported by Edmiston and coworkers [50–54]. The first step involved mixing two solutions, bis(trimethoxysilyl) ethyl benzene (BTEB) dissolved in acetone and tetrabutylammonium fluoride dissolved water. Mixing the two solutions initiated gelation and the resulting gel was kept at 25 °C for 6 days. The gel was then rinsed with acetone, ground and derivatized using 5 v/v% solution of chlorotrimethylsilane in acetonitrile for 24 h. Finally, the solution was filtered and the remaining solids were washed with acetone and dried at 110 °C to obtain SOMS, as a white color solid.

The high-temperature treated SOMS, referred to as H-SOMS in this paper, was prepared by calcining SOMS under inert conditions. The SOMS was first saturated with acetone until no more volume expansion was observed (fully swollen). This was to increase the surface area and pore volume of the SOMS to its swollen state before the calcination was carried out. The saturated SOMS was placed into a horizontal quartz tube which was flushed with nitrogen for 5 min to remove oxygen inside. The quartz tube was then transferred into the center of a heated furnace at 600 °C while nitrogen was flowing through the tube. SOMS was calcined for 2 h under a nitrogen flow. Afterwards, the calcined SOMS, i.e. H-SOMS, was obtained as a brown-colored solid.

Incipient wetness impregnation (IWI) technique was used to create Pd particles within the SOMS support. A known amount of

Pd acetate solid (1% wt Pd relative to the support) was dissolved in sufficient acetone to achieve a solution and aid the swelling process. These precursors were chosen to allow the opening of the pores during impregnation. The Pd precursor solution was then added drop-wise until SOMS was saturated. The Pd impregnated SOMS was then dried at RT with constant agitation. The entire process was repeated multiple times to use all of the precursor solution. The 1% Pd/SOMS was obtained after the final drying step. For H-SOMS, similar preparation steps were employed, except that the solids were dried at 100 °C. The Pd/SOMS was dried at RT to make sure that the sample did not go through a thermal treatment.

The reduction of Pd (Pd²⁺ to Pd⁰) was carried out chemically using a saturated solution of NaBH₄ in 95% ethanol in the case of 1% Pd/SOMS. After Pd/SOMS was soaked in the NaBH₄ solution, it was filtered and rinsed with ethanol followed by deionized water. The remaining solids were dried at 60 °C. For the reduction of 1% Pd/H-SOMS, 5% H₂/He gas was utilized at which the catalyst was reduced *in-situ* at 400 °C for 1 h. The reduced Pd/H-SOMS was then treated with helium at 450 °C for 30 min for surface clean-up. The reason for using two different pre-reduction treatments was due to thermal decomposition of SOMS. As reported in the TPD results (Section 3.1) SOMS started to decompose around 400 °C. Therefore, if Pd/SOMS was pre-reduced using the hydrogen at 400 °C, chemical properties of SOMS would change. Thus, an alternative way was implemented (NaBH₄) which can be conducted at 25 °C.

2.2. Catalyst characterization

2.2.1. Temperature programmed decomposition (TPD)

TPD experiment was performed using a Carbolite MTF 10/15/130 furnace where a 4 mm-ID quartz reactor was located at the center of the furnace. SOMS was first placed inside the quartz reactor supported by quartz wool plugs. A continuous flow of pure helium was then introduced to the reactor at 30 ml/min at RT and the effluent was sent to a mass spectrometer (MS – MKS-Cirrus II) operated in scanning ion mode. Once a stable MS signal of helium was obtained, the furnace temperature was increased gradually at a ramp rate of 10 °C/min from RT to 800 °C and stayed there for 150 min. For identification of the products, signals from *m/z* = 2 to 60 were monitored.

2.2.2. Surface area, pore size and pore volume

Brunauer-Emmet-Teller (BET) surface area and Brunauer-Joyner-Helena (BJH) pore volumes of SOMS, H-SOMS, Pd/SOMS and Pd/H-SOMS were measured using Micromeritics ASAP 2020 (accelerated surface area and porosimetry) instrument. The samples were first degassed for 12 h at 130 °C under 2 μm Hg of vacuum and transferred to the analysis port to obtain nitrogen adsorption/desorption isotherms of the samples. The analysis was conducted at liquid nitrogen temperature and the saturation pressure of nitrogen was measured frequently to improve accuracy of the relative pressure calculation. The BJH pore size distribution was acquired using the desorption branch of the isotherm.

2.2.3. High angle annular dark field (HAADF) scanning transmission electron microscopy (STEM)

Identification and particle size information of Pd were obtained by taking HAADF-STEM images. A small amount of Pd/SOMS and Pd/H-SOMS was first dispersed in ethanol and stirred rigorously for 5 min using a sonicator. The resulting solution was then placed drop-wise on a Tedpella, Inc. 200 mesh copper grid coated with lacey carbon sample holder. After a drying step at RT for 1 h, the sample holder was transferred to FEI Tecnai F20 instrument equipped with a high brightness field emission electron gun (FEG) operated at 200 kV and a HAADF detector. Multiple images were

taken in different spots of the TEM grid to ensure uniformity of the Pd particle sizes.

2.2.4. ²⁹Si nuclear magnetic resonance (NMR) and infrared (IR) spectroscopy

²⁹Si MAS (Magic Angle Spinning) NMR spectra were collected over Pd/SOMS and Pd/H-SOMS samples using Bruker DPX 300 MHz NMR instrument at a resonance frequency of 59.6 MHz. Tetramethylsilane (TMS) was used as a reference for 0 ppm. SpinWorks 4.2.3.0 ©2016 software was utilized for line broadening of the spectra. Infrared spectra of SOMS and H-SOMS were acquired using a Thermo Nicolet 6700 FTIR Spectrometer with a diamond attenuated total reflectance (ATR) accessory. The material was ground to a particle size <250 μm. Spectra were acquired at 4 cm⁻¹ resolution averaging 16 scans.

2.2.5. Static vapor-phase adsorption

Pyris 1 Perkin-Elmer thermogravimetric analyzer (TGA) was used to obtain acetone adsorption profiles of the Pd/SOMS and Pd/H-SOMS catalysts at *T* = 25 °C. The sample was placed inside a sealed container in which a saturated acetone vapor was introduced. The mass of the sample was first measured at 0 s and collected continuously every 6 s until no further mass increase was observed. Percentage increase in mass (%) was calculated for each sample.

2.2.6. CO chemisorption

Carbon monoxide (CO) was used as a probe molecule in order to obtain the dispersion of Pd particles on SOMS and H-SOMS supports. A similar reactor system used for the TPD experiment was employed with an addition of a 6-port valve which allowed introduction of CO gas in pulses. The samples were pre-reduced prior to CO chemisorption. Pd/SOMS was reduced chemically using NaBH₄ solution while Pd/H-SOMS was pre-reduced *in-situ* under 5% H₂/He at 400 °C. The CO chemisorption experiments were conducted at RT sending known amounts of CO to the catalyst bed in pulses until saturation was reached. The outlet of the reactor was connected to a MS and the amount of CO uptake at each pulse was quantified based on the feed (250 μl sample loop) which was calibrated before the reaction. The Pd dispersion was calculated as “number of moles of Pd reacted with CO/number of moles of Pd in the sample”. The calculation was performed assuming a 1:1 ratio of Pd to CO.

2.2.7. In situ X-ray absorption near edge structure (XANES) spectroscopy

In situ XANES spectra were collected at the insertion device (Sector 10-ID) beamline of the Materials Research Collaborative Access Team (MRCAT) located at the Advanced Photon Source (APS) – Argonne National Laboratory. The measurements were performed in transmission mode using a Si (111) monochromator with a harmonic rejection mirror covered with Pt and Rh. The 1% Pd/SOMS and 1% Pd/H-SOMS samples were first finely ground with/without addition of silica as a diluent to obtain a proper Pd K edge (24.36 KeV) jump. The prepared samples were then pressed gently into a six-well sample holder made of stainless steel which was inserted horizontally to a 45 cm long quartz tube (2 cm ID). The inlet and outlet of the quartz tube were connected to stainless steel tubing fitted with Swagelok fittings to allow *in-situ* experiments.

All the samples went through the same treatment conditions as follows: (1) helium treatment at RT (2) oxidation at 250 °C, (3) reduction at 150 and 250 °C, (4) re-oxidation at 150 and 250 °C. The sample was kept at each temperature for 30 min which corresponds to 3 h of experiment time. The gases used for oxidation and reduction were 5% O₂/He and 3.6% H₂/He, respectively. The switch between gases was made after the existing gas in the lines was flushed with helium. Lastly, the quantification of Pd oxidation

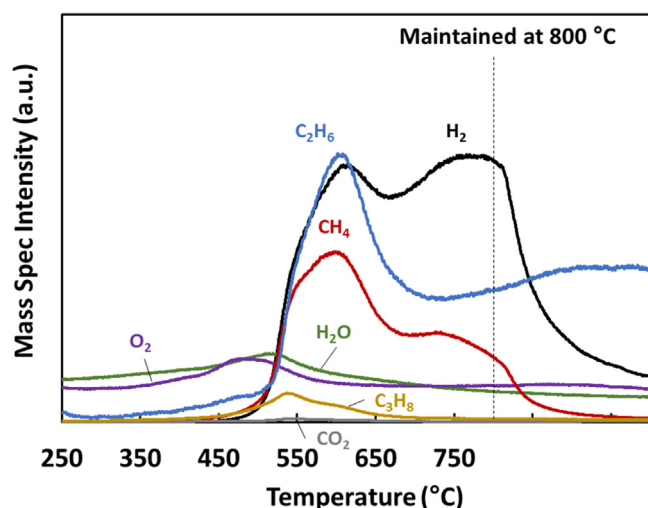


Fig. 1. Evolved species during temperature programmed decomposition of SOMS under helium.

states was obtained using WinXAS software by linear combination between the spectra of Pd metal foil and PdO standard.

2.3. Catalyst activity testing

Gas-phase HDC of TCE was performed over 1% Pd/SOMS and 1% Pd/H-SOMS catalysts in a bench-scale fixed bed flow reactor system. A nitrogen stream was saturated with TCE using a bubbler at 35 °C. The N₂ stream saturated with TCE was then diluted with additional nitrogen in order to lower the concentration of TCE. Afterwards, hydrogen was added to the mixture resulting in a feed composition of 0.7% (by volume) TCE in excess hydrogen at a TCE-to-hydrogen molar ratio of 1:30. For HDC of TCE experiments in the presence of water vapor, the water concentration in the feed was 4%. The reactants were then sent to the catalyst bed located inside the 4 mm-ID quartz reactor. 50 mg of catalyst was used with a 30 ml/min of total reactant flow. The catalyst bed height was 6 mm. The gas hourly space velocity was obtained as 24,000 h⁻¹. The gas-phase HDC of reaction was carried out at 100, 150 and 200 °C. The reactor temperature was increased/decreased by an Omega CSC232 PID temperature controller which adjusts the temperature of the surrounding furnace (Carbolite, MTF 10/15/130). All the catalysts were pre-reduced prior to the reaction (Pd/SOMS – NaBH₄ and Pd/H-SOMS – *in situ* H₂). For accurate comparison, the catalytic activities of the samples were obtained at equal gas hour space velocity (GHSV). The feed and the reactor effluent were analyzed using an online gas chromatograph (Shimadzu Scientific 2010) with a Q-bond column and a flame ionization detector (FID). The oven temperature, injector pressure/linear velocity of the He carrier gas were altered to obtain proper separation of the product peaks including TCE, partially dechlorinated compounds, short-chain hydrocarbons, especially the major product ethane.

3. Results and discussion

3.1. Temperature-programmed decomposition over SOMS

The procedure for the preparation for H-SOMS was developed after observing the results of a temperature-programmed decomposition (TPD) experiment conducted under helium environment to identify the species that were evolved. Fig. 1 shows the profiles for different mass-to-charge ratios (*m/z*) of 2 (H₂), 15 (CH₄), 18 (H₂O), 28 (C₂H₆), 29 (C₃H₈), 32 (O₂) and 44 (CO₂). The *m/z* values were selected considering the fragmentations of molecular ions of

Table 1

BET surface area, pore volume and average pore diameter of SOMS, H-SOMS, 1% Pd/SOMS and 1% Pd/H-SOMS.

| | BET Surface Area (m ² /g) | Pore Volume (cm ³ /g) | Averaged Pore Diameter (Å) |
|--------------|---|-------------------------------------|-------------------------------|
| SOMS | 493 | 1.03 | 84 |
| H-SOMS | 682 | 2.07 | 122 |
| 1% Pd/SOMS | 479 | 0.56 | 47 |
| 1% Pd/H-SOMS | 546 | 1.09 | 75 |

the gases so that there are no contributions in the intensities to each other. As it is demonstrated in the TPD results, the decomposition of SOMS started around 450 °C releasing short chain hydrocarbons such as methane, ethane and propane as well as hydrogen, oxygen and water. The fact that majority of the products are methane and ethane indicates the C–C bond cleavages of aromatic rings within the SOMS structure. The color of the material turned black after calcination, possibly indicating carbon deposition on the SOMS surface. After the reactor temperature was increased to 800 °C, the temperature was maintained for 2 h. By then, most of the gases were liberated from the SOMS material at which the *m/z* signals returned to its original intensities.

The absence of oxygen in the atmosphere led to trace amounts of CO₂ formation, although slight increase in CO₂ intensity was observed which can be due to remaining surface oxygen. It is important to note that the synthesis of H-SOMS is not an oxidation process; therefore, the carbon-containing organic groups are still present in the SOMS matrix.

3.2. Surface area, pore volume and average pore diameter

Table 1 shows the BET surface area, pore volume and averaged pore diameter of the samples used for this study. It was observed that the BET surface area of SOMS was increased significantly after high-temperature treatment was performed (493 m²/g → 682 m²/g). Likewise, the increase in pore volume and average pore diameter possibly indicates formation of larger pores during the calcination process. The surface area of Pd/H-SOMS is higher because H-SOMS was synthesized by exposing the SOMS support to a “thermal shock” treatment when acetone was fully adsorbed inside the SOMS structure. As mentioned previously, the swelling increases the internal surface area and pore volume. Therefore, when it is calcined, it retains the open-pore structure of the SOMS (H-SOMS) support obtaining higher surface area and pore volume compared to SOMS. The impregnation of Pd particles on SOMS and H-SOMS supports led to a decrease in BET surface area, pore volume and average pore diameter compared to its bare SOMS and H-SOMS supports, suggesting partial closure (or filling) of the pores through the impregnation process.

It should be pointed out that the N₂ physisorption measurements do not show the real surface area and pore volume of swollen SOMS. The N₂ adsorption on the SOMS support does not take place inside the swollen-SOMS structure; it only probes the unswollen surface of the SOMS. Due to the experimental requirements of N₂ physisorption experiments such as the low-pressure and temperature during N₂ adsorption and desorption, SOMS is present in the unswollen state.

3.3. Pd particle size

The Pd particle sizes of 1% Pd/H-SOMS and 1% Pd/SOMS samples were measured using STEM technique (Fig. 2). The particle size distribution obtained from counting 50–100 particles from the images are shown in Fig. 3. With respect to Pd/SOMS, Pd particles (shown as white dots) on SOMS support are not well distributed compared to Pd/H-SOMS catalyst. The average Pd particle size for

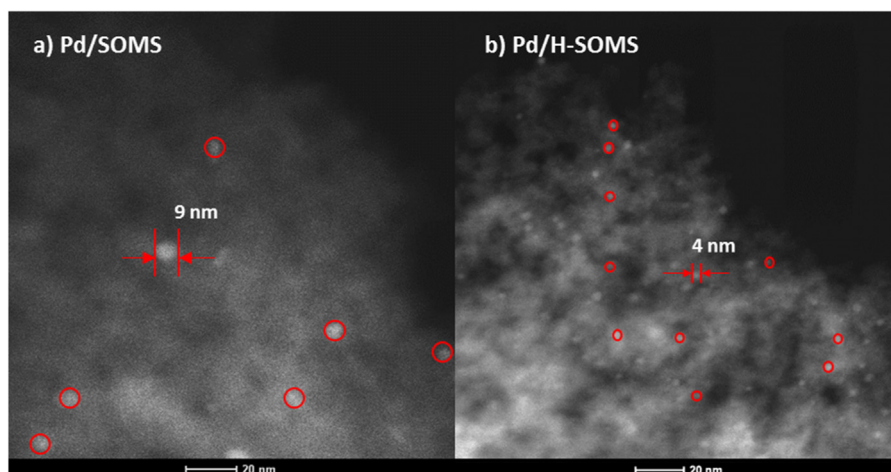
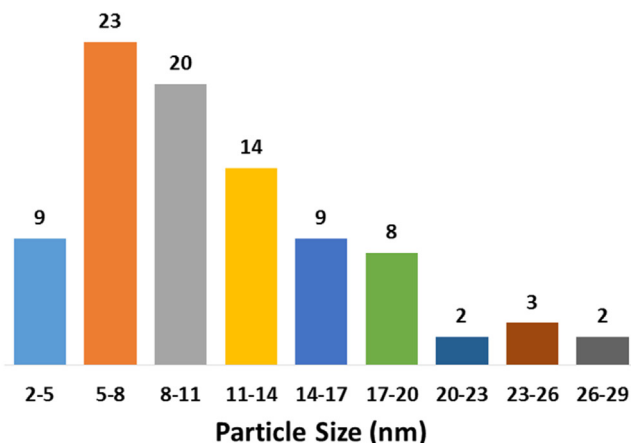


Fig. 2. STEM images of Pd/SOMS and Pd/H-SOMS (white dots represents Pd particles).

(a) Pd/SOMS

Average: 11 nm
SD: 6 nm



(b) Pd/H-SOMS

Average: 2.8 nm
SD: 0.5 nm

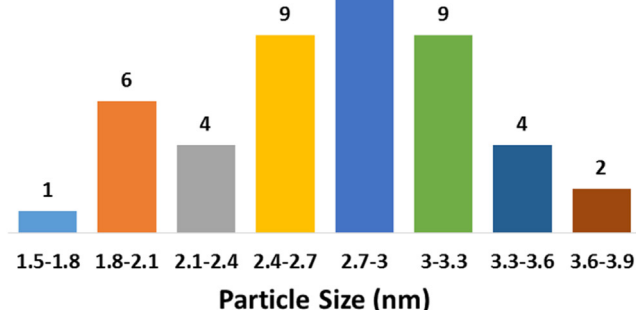


Fig. 3. Particle Size Distribution for Pd/SOMS and Pd/H-SOMS.

Pd/SOMS was estimated around 11 nm with standard deviation (SD) of 6 nm. Also, larger Pd particles above 20 nm were observed (not shown in the figure). The non-uniform distribution of Pd particles and formation of larger Pd clusters are most likely attributed to the animated nature of the SOMS matrix during the Pd impregna-

tion, drying and/or NaBH_4 chemical reduction processes. Especially, while SOMS returns to its tensioned matrix due to evaporation of the organic solvents (acetone), it may cause relocation and agglomeration of the Pd particles.

On the other hand, the Pd particles on the H-SOMS support were smaller than on SOMS with an average Pd particle size of 3 nm and SD of 0.5 nm. The Pd particles were more homogeneously dispersed on H-SOMS with a narrower particle size distribution compared to Pd/SOMS catalyst. As it is seen Fig. 2 (a) and (b), which are on the same scale, significantly more Pd particles with a smaller particle size can be observed for Pd/H-SOMS. It is worth noting that the process of Pd impregnation on H-SOMS does not accompany with any physical expansion/compression (swelling) of the support matrix; therefore, it can be speculated that the swelling behavior plays an important role in determination of the Pd particle size.

3.4. Chemical structure of SOMS and H-SOMS

3.4.1. Infra-red spectroscopy

IR spectra acquired over SOMS and H-SOMS were compared to understand the change in structural properties of the material due to heat treatment. Fig. 4 (a) shows the spectra collected in the range of 600–1800 cm^{-1} and (b) from 1800 to 3600 cm^{-1} . In the spectrum for SOMS, the two bands located at 1034 and 1093 cm^{-1} are correlated to the Si–O–Si stretching [55]. The 1093 cm^{-1} peak is assigned to longitudinal-optic (LO)-transverse-optical (TO) splitting with an additional peak at 1188 cm^{-1} [56], specifically, indicating LO_3 - TO_3 antisymmetric stretching, which was reported due to long range coupling of Coulomb interactions [57–59]. These bands are not present in the H-SOMS spectra and are replaced with a broad Si-O-Si stretching peak characteristic of general silica materials. Volumetric expansion of H-SOMS was found to be essentially eliminated when tested using exposure to neat acetone. Therefore, the antisymmetric stretching LO_3 (1188)- TO_3 (1093) cm^{-1} vibration appears to be related to the swelling capability of the SOMS material.

Near the lower wavelength region of the SOMS spectrum, bands at 688 and 754 cm^{-1} were likely to be assigned to out-of-plane C–H bending/wagging vibrations of mono and meta substituted benzene compounds [55,60]. Similarly, for 1400–1700 cm^{-1} region, several peaks were appeared which were linked to in-plane C–H bending of substituted benzenes [60]. These bands are not observed for H-SOMS which leads to a conclusion that almost all of the aromatic rings were decomposed due to the high temperature treatment.

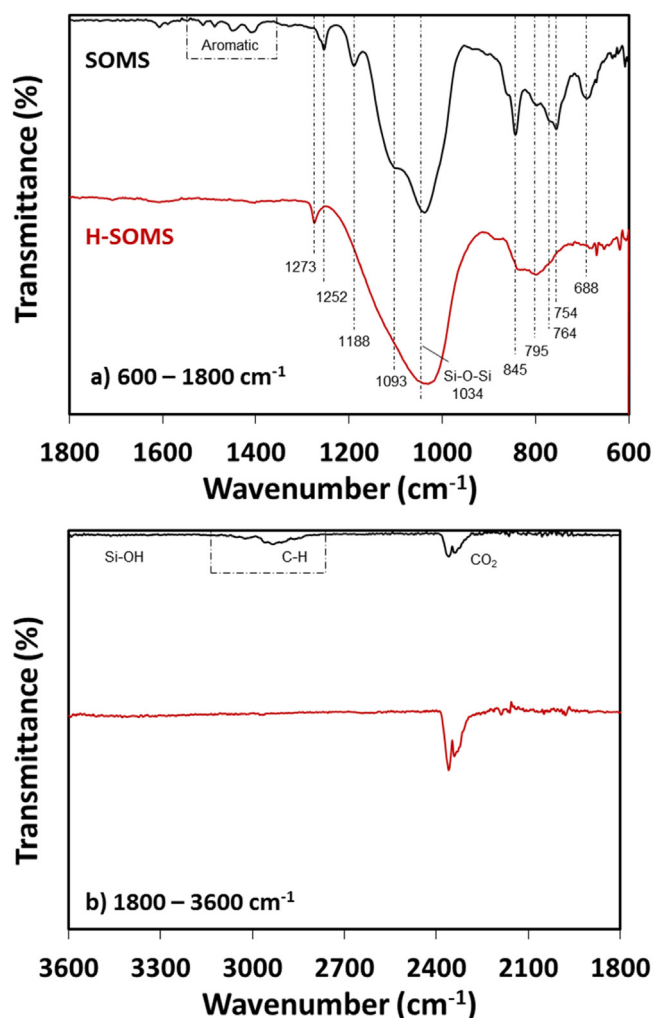


Fig. 4. IR spectra for SOMS and H-SOMS in the range of (a) 600–1800 cm⁻¹ and (b) 1800–3600 cm⁻¹.

The IR bands at 764, 845 and 1252 cm⁻¹ of the SOMS spectrum indicate that O–Si–(CH₃)₃ are present in the SOMS material [55]. The formation of trimethylsilyl group due to the derivatization of residual Si–OH during SOMS synthesis (chlorotrimethylsilane treatment) [51]. This end-caps the residual Si–OH groups which help to prevent additional Si–O–Si bond formation upon drying the gel, preserving flexibility of the SOMS structure. However, the intensities of the 763 and 1252 cm⁻¹ bands significantly decrease for H-SOMS; instead, two bands of 795 and 1273 cm⁻¹ appear in the spectrum. The shifts of 763–795 cm⁻¹ and 1252–1273 cm⁻¹ is relevant to generation of dimethyl groups (O–Si–(CH₃)₂–) [55]. This shows that the calcination of SOMS under inert conditions causes the partial destruction of the end-capped trimethylsilyl and forms additional linkages of the polysiloxane network. Si–OH bonds were not observed for neither SOMS nor H-SOMS samples. Formation of additional Si–O–Si crosslinks decreases the internal flexibility and eliminates the swelling motif.

3.4.2. ²⁹Si NMR

The structure characteristics of SOMS and H-SOMS were investigated in detail using ²⁹Si NMR. Fig. 5 shows the spectra collected over the Pd/SOMS and Pd/H-SOMS materials. The NMR spectrum of Pd/SOMS consists of three distinctive peaks, one at 8 ppm and the other two at –59 ppm and –68 ppm. The region around 8 ppm corresponds to the (SiO)₄–Si* where the center Si* is attached to a SiO group and three substituents. [61] In this case, all three

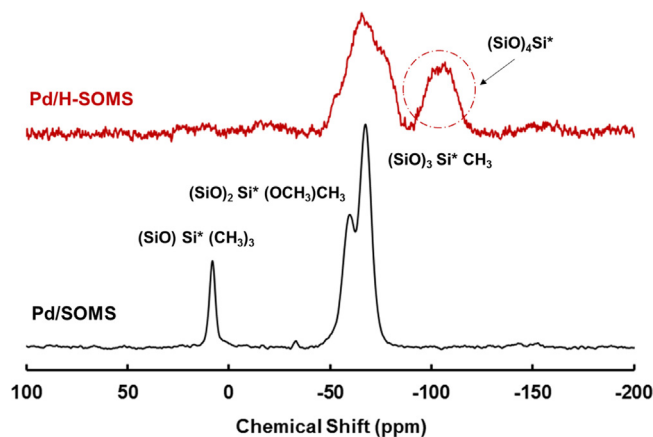


Fig. 5. NMR spectra for Pd/SOMS and Pd/H-SOMS.

substituents are methyl groups of (SiO)–Si*–(CH₃)₃, showing agreement with the IR data [61–63]. The peak at –59 and –68 ppm are due to (SiO)₂–Si*–(OCH₃)R⁶³ and (SiO)₃–Si*–R⁶⁴ chemical sites where Si* is bonded to three oxygens (O) [65]. Specifically, it is likely that the substituted benzenes may contribute to the intensity of these two peaks [66].

The spectrum for H-SOMS shows the high temperature treatment to have a significant impact on the chemical structure of SOMS. The appearance of peak around –105 ppm region indicates (SiO)₄–Si* at which Si* is linked to four SiO groups [64]. The formation of additional Si–O bonds is due to decomposition of SOMS releasing mostly hydrocarbons in the gas phase, but not oxygen-containing species. It can be concluded that the thermal treatment process under inert conditions transferred SOMS into a more silica-like material (SiO₂), however, it should be noted that the peak at around –65 ppm for H-SOMS still indicates the presence of organic species in the H-SOMS support matrix.

3.5. Acetone vapor adsorption in static conditions

The adsorption capacity and affinity of 1% Pd/SOMS and 1% Pd/H-SOMS catalysts for acetone were examined in static conditions. In Fig. 6, the percent increase in the mass of the two samples are plotted with respect to the exposure time to the acetone vapor. The purpose of this experiment is to confirm that the Pd/H-SOMS still has affinity towards organic molecules and preserves its hydrophobicity even though it went through a thermal shock treatment and a calcination process. Acetone vapor was used as a probe molecule since it is shown earlier that SOMS adsorbs acetone the most compared to other organic molecules including TCE [49]. Pd/SOMS shows a high adsorption capacity for acetone and is highly hydrophobic (Fig. 6 (a)). At 120 min of adsorption time, the mass of the Pd/SOMS increased twice to its original value. For the Pd/H-SOMS sample, the adsorption capacity for acetone was decreased compared to Pd/SOMS (Fig. 6 (b)). As mentioned earlier, the calcination of SOMS results in reduction of swelling capability and hydrophobicity. The swelling of the SOMS structure is closely associated with the increase in pore volume and surface area, which increase the adsorption capacity of the material. Therefore, these results are expected since the extent of swelling for Pd/H-SOMS is much less than Pd/SOMS catalyst. However, as it can be seen, Pd/H-SOMS still has a higher affinity for acetone than it does for water, showing that it retains some hydrophobicity.

Fig. 6 (c) compares the first five minutes of the adsorption process for both materials. The rate of adsorption of acetone is seen to be significantly faster in Pd/H-SOMS than Pd/SOMS. This result is important in showing that although the adsorption capacity of the

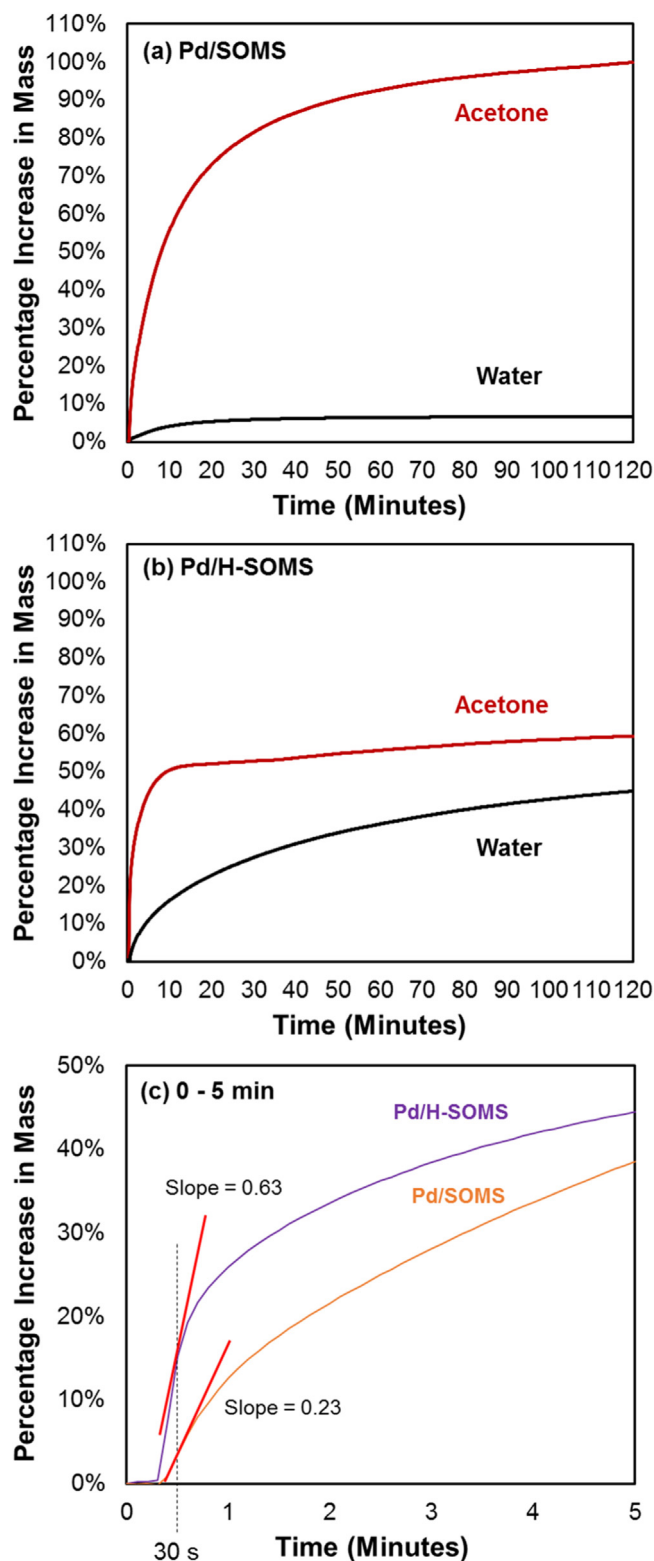


Fig. 6. Absorption of acetone and water over (a) Pd/SOMS (b) Pd/H-SOMS, (c) comparison of absorption of acetone over Pd/SOMS and Pd/H-SOMS in the first 5 min.

material has decreased upon heat treatment, its rate of adsorption has increased, presumably by creating a more open pore structure.

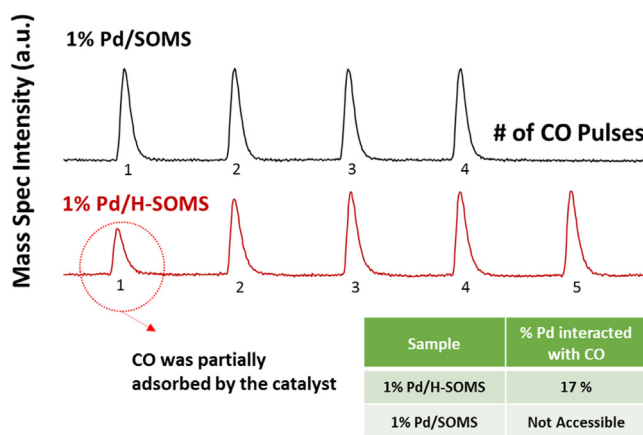


Fig. 7. Results of CO Chemisorption.

3.6. Accessibility of Pd to CO, O₂ and H₂

3.6.1. CO chemisorption results

CO chemisorption technique has been used in many studies in order to investigate the Pd dispersion on a catalyst surface [67–69]. In this study, this technique was used to probe the site accessibility of the Pd particles inside the pores by introducing known volumes of CO to the catalyst bed using pulses. Both catalysts were pre-reduced prior CO chemisorption, either chemically (Pd/SOMS) or thermally using hydrogen (Pd/H-SOMS). Fig. 7 shows the MS intensity for CO signal ($m/z=28$) where each peak corresponds to a single pulse. As can be seen, the CO molecules in the first pulse were partially adsorbed by the 1% Pd/H-SOMS, specifically by the reduced Pd sites of the catalyst, whereas no adsorption of CO was observed for 1% Pd/SOMS. This is due to the inaccessibility of Pd sites to the incoming CO molecules in case of the Pd/SOMS since in its unswollen state, the pores are closed and the Pd sites are sequestered inside the pores. In Pd/H-SOMS, on the other hand, the pores are open and the Pd sites are accessible, leading to CO adsorption.

As the CO pulses were continued, the intensity of the CO peaks leveled off to the original value of the feed for Pd/H-SOMS. The % Pd sites accessed by CO was calculated as ~17% based on a ratio of 1:1 for Pd and CO [70,71].

3.6.2. In-situ XANES during oxidation, reduction and re-oxidation

The oxidation state of Pd was followed by XANES while the catalyst was under oxidizing or reducing environments. The switches from O₂ to H₂ and H₂ to O₂ were made at 30 min and 90 min, respectively. The XANES spectra collected at 90 min and 150 min are shown in Fig. 8. Spectra for references, PdO and metallic Pd are also shown in the same figure. The calculated oxidation states are shown in Fig. 9. Both samples start out with an average oxidation state of 0.94 and 0.77 for Pd/SOMS and Pd/H-SOMS, respectively. At 250 °C during the 30 min of oxidation, a sudden rise in Pd oxidation state in Pd/H-SOMS was observed whereas no oxidation of Pd was seen over the Pd/SOMS catalyst. This can be attributed to the presence of severe mass transfer limitations for oxygen molecules to enter the catalyst pores in the SOMS support structure. In other words, it shows that most of the Pd particles are embedded inside the SOMS matrix thereby being not accessible to the incoming molecules unless they are organics. The difference is ascribed to the open-pores of the H-SOMS chemical structure where majority of the Pd particles are readily accessible.

The reduction was performed subsequently using 5% H₂ at 150 °C. After 15 min of reduction, all the samples were significantly reduced reaching an oxidation state of 0.20 and 0.15 for Pd/SOMS

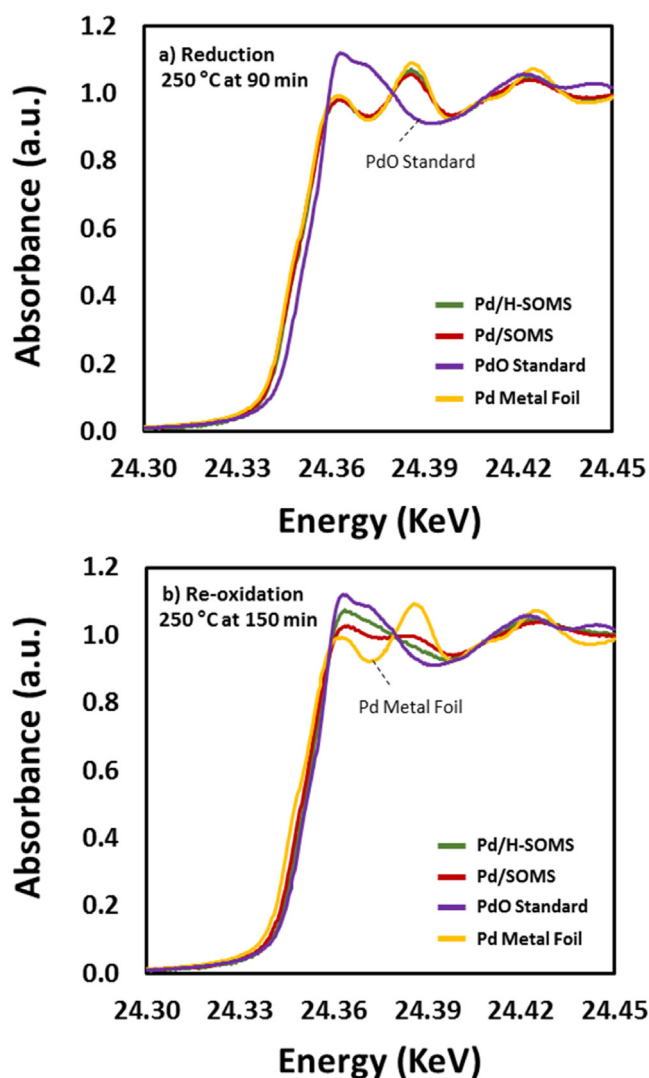


Fig. 8. XANES spectra collected under reducing and oxidizing reaction conditions.

and Pd/H-SOMS, respectively. Likewise, at 250 °C, the extent of reduction of Pd/H-SOMS and Pd/SOMS were maintained close to Pd⁰. The reduction of Pd of Pd/SOMS catalyst is interesting, especially considering the fact that oxygen was not able to access the

Pd sites. Therefore, it is probable that the adsorption takes place depending on the size of the molecule, i.e. hydrogen molecule is much smaller than oxygen. The SOMS support acts as a molecular sieve where larger molecules are not allowed to pass through to get in contact with the Pd particles located inside the SOMS matrix. However, it should be pointed out that this is only valid when Pd/SOMS is in unswollen state. Regardless of the molecular size, Pd sites are accessible to the reactants if the SOMS is in the swollen state.

Re-oxidation of the samples was carried out following the reduction treatment. Again, similar to the conclusion made previously, Pd supported on SOMS remained in a more reduced state compared Pd supported on H-SOMS.

3.7. Gas-phase HDC of TCE

3.7.1. Catalytic activity

The catalytic performances of 1% Pd/SOMS and 1% Pd/H-SOMS were investigated for gas-phase HDC of TCE. The reaction was performed in a continuous gas-phase fixed bed reactor system at reaction temperatures of 100, 150 and 200 °C. It is expected that there will be an influence of the reaction temperatures on the textural and morphological characteristics of Pd/SOMS. Pd/H-SOMS, on the other hand, is thermally more stable and less sensitive since the support SOMS was treated at 600 °C before Pd impregnation. Although no characterization was performed on the post-reaction catalysts, effect of the reaction conditions on the material characteristics of Pd/H-SOMS are expected to be smaller than that on Pd/SOMS.

The TCE conversions obtained over the two samples are shown in Table 2. As expected, 1% Pd/SOMS was significantly less active for gas-phase HDC of TCE reaction in the entire temperature range compared to the 1% Pd/H-SOMS catalyst. This is attributed to the lower degree of swelling of the Pd/SOMS sample in the gas-phase compared to liquid-phase, which results in a decrease in the number of Pd active sites available for the HDC of TCE. It should be noted that the actual number of accessible metallic Pd sites over Pd/SOMS during the HDC reaction is certainly higher since, during the approach to steady-state, SOMS support swells, opening the pores, as seen in our previous work [49]. The adsorption of TCE increases the availability of the Pd metal in the SOMS matrix by opening up the catalyst pores while physical expansion of the Pd/SOMS catalyst proceeds. However, for gas-phase HDC of TCE, a significantly low catalytic performance of 1% Pd/SOMS was obtained. Thus, whether the pores are opened enough so that Pd

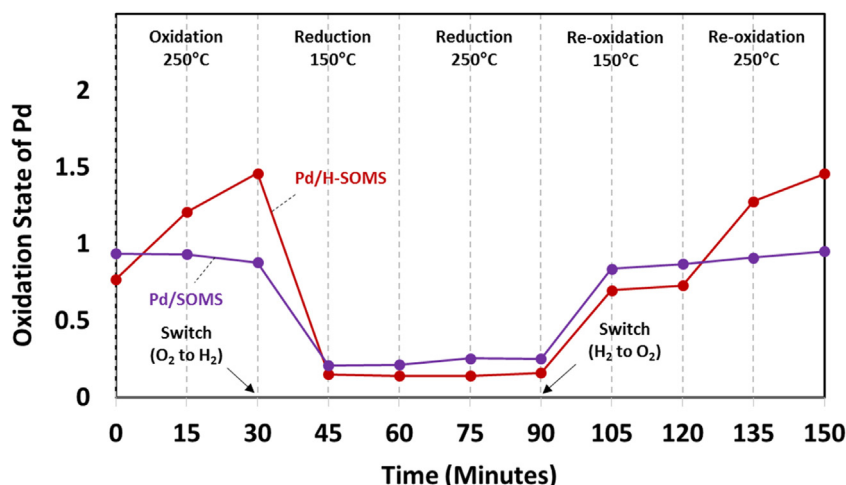


Fig. 9. Change in oxidation states of Pd for Pd/SOMS and Pd/H-SOMS in reducing and oxidizing reaction conditions.

active sites are all accessible for the HDC reaction in the gas-phase remains as a question. In addition, the fact that an opposite trend was obtained in liquid-phase HDC of TCE where the 1% Pd/SOMS catalyst was more active than 1% Pd/Al₂O₃ (data not shown) supports the statement that the swelling, i.e. expansion of the volume of SOMS, is somewhat limited in the gas phase.

When Pd was impregnated on calcined SOMS under inert conditions (Pd/H-SOMS), a significant improvement of the catalytic activity was observed. This is in good agreement with the CO chemisorption results where more Pd metallic sites were shown to be accessible. Thus, an increased conversion level from Pd/H-SOMS is expected, since under gas-phase conditions at elevated temperature the amount of adsorption of TCE is insufficient to open the pores of Pd/SOMS to allow access to “sequestered” catalytic sites. At 200 °C, the TCE conversion obtained over the Pd/H-SOMS catalyst was 92.5% whereas only 38.3% was obtained over Pd/SOMS.

The main products of HDC of TCE is hydrogen chloride (HCl) and ethane (C₂H₆). As shown in Table 2, Pd/H-SOMS showed a higher HDC percentage and higher ethane selectivity. Among the by-products, partially dechlorinated products such as *cis*-1,2 DCE and 1,1 DCE gave higher yields compared to *trans*-1,2 DCE, especially when lower TCE conversions were obtained. Ethylene was acquired in large amounts over the Pd/SOMS at higher temperatures. Also, vinyl chloride and even number of higher-carbon products such as *n*-butane were observed [23,72]. Significant catalyst coking was observed at a reaction temperature of 200 °C for Pd/H-SOMS catalysts, indicating that Pd sites have both C–Cl and C–C bond cleavage activity.

Table 2
Gas-phase catalytic activity data of Pd/SOMS and Pd/H-SOMS for HDC of TCE.

| | Pd/SOMS | | | Pd/H-SOMS | | |
|-----------------------|---------|--------|--------|-----------|--------|--------|
| | 100 °C | 150 °C | 200 °C | 100 °C | 150 °C | 200 °C |
| % HDC | 4.2 | 7.7 | 25.6 | 5.2 | 24 | 87.8 |
| % TCE Conv. | 9.3 | 13.7 | 38.3 | 10.2 | 34.8 | 92.5 |
| Selectivity (%) | | | | | | |
| Ethane | 9.0 | 12.3 | 18.6 | 19.6 | 44.2 | 71.4 |
| Ethylene | 4.3 | 15.2 | 29 | 3.9 | 6.9 | 15.5 |
| Vinyl Chloride | 5.3 | 10.9 | 5.0 | 3.9 | 2.2 | 0.6 |
| <i>n</i> -butane | 4.3 | 0.7 | 1.0 | 4.9 | 4.9 | 3.1 |
| 1,1 DCE | 31.0 | 26.8 | 19.3 | 18.7 | 15.5 | 4.3 |
| <i>trans</i> -1,2 DCE | 3.3 | 1.6 | 1.2 | 9.8 | 4.0 | 2.1 |
| <i>cis</i> -1,2 DCE | 42.8 | 32.6 | 25.9 | 39.2 | 22.3 | 2.9 |

Table 3

Effect of addition of water in the reactant stream on catalytic activity of Pd/SOMS and Pd/H-SOMS.

| | Pd/SOMS | | | Pd/H-SOMS | | |
|-----------------------|---------|--------|--------|-----------|--------|--------|
| | 100 °C | 150 °C | 200 °C | 100 °C | 150 °C | 200 °C |
| % HDC | 5.0 | 9.1 | 22.8 | 5.6 | 29.3 | 92.1 |
| % TCE Conv. | 10.1 | 16.6 | 35.3 | 11.1 | 40.9 | 96.6 |
| Selectivity (%) | | | | | | |
| Ethane | 11.9 | 12 | 15.9 | 23.4 | 48.4 | 85.6 |
| Ethylene | 6.9 | 14.5 | 25.8 | 0.9 | 5.9 | 3.3 |
| Vinyl Chloride | 6.9 | 9.0 | 4.2 | 1.8 | 2.7 | 0.3 |
| <i>n</i> -butane | 5.0 | 6.0 | 11.6 | 3.6 | 6.6 | 5.9 |
| 1,1 DCE | 21.8 | 22.3 | 17.6 | 21.6 | 13.2 | 1.4 |
| <i>trans</i> -1,2 DCE | 7.9 | 7.2 | 0.8 | 5.4 | 3.9 | 2.1 |
| <i>cis</i> -1,2 DCE | 39.6 | 28.9 | 24.1 | 43.2 | 19.3 | 1.4 |

3.7.2. Effect of addition of water

The effect of the addition of water vapor to the reactant stream on catalytic activities of 1% Pd/SOMS and 1% Pd/H-SOMS was also examined (Table 3). The addition of water to the reactant mixture did not affect the catalytic performance of the Pd/H-SOMS or Pd/SOMS catalysts. The TCE conversion levels in the absence and presence of water were very similar (Fig. 10). This is mainly attributed to the strong hydrophobicity of SOMS and H-SOMS supports that repels the water molecules thereby restricting the adsorption of water on the catalyst surface. Even though, the degree of hydrophobicity of the H-SOMS was somewhat reduced compared to SOMS, the results show that the inhibition effect of water on the catalytic activity of Pd/H-SOMS is negligible. However, at 200 °C, as observed before, carbon deposition on the Pd/H-SOMS catalyst was encountered.

4. Conclusions

The catalytic performance of Pd supported on SOMS and high-temperature treated SOMS (H-SOMS) was investigated for the gas-phase HDC of TCE. In our previous results, it was reported that 1% Pd/SOMS has shown significantly better catalytic performance compared to 1% Pd/Al₂O₃ catalyst under liquid-phase HDC of TCE conditions. This was attributed to the swelling capability of SOMS which concentrated the TCE reactants in the vicinity of the Pd active sites. However, in the gas-phase, the catalyst pores of Pd/SOMS were not fully open, thereby leading to lower catalytic of Pd/SOMS than Pd/Al₂O₃. Thus, in this study, a sample of SOMS saturated with

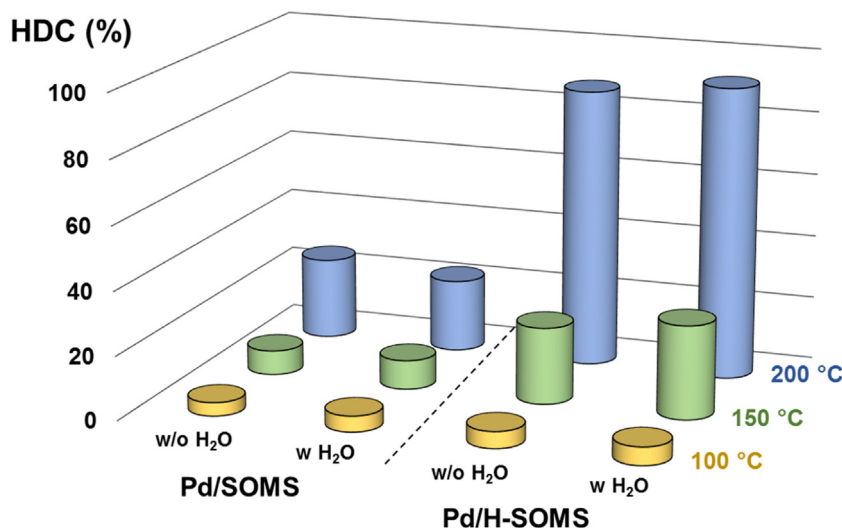


Fig. 10. Comparison of percentage HDC of TCE in the presence/absence of water in the reactant stream for Pd/SOMS and Pd/H-SOMS catalysts.

acetone was calcined under inert conditions at high temperature in order to retain its swollen state resulting better pore accessibility. The surface analysis results exhibited higher BET surface area, pore volume and averaged pore diameter over the H-SOMS material compared to SOMS. After Pd was placed on both supports, the Pd dispersion of Pd/H-SOMS was greatly improved than that of Pd/SOMS based on the STEM images. Also, the average Pd particle size was smaller for Pd/H-SOMS. XANES results exhibited that Pd sites on H-SOMS were accessible to both H₂ and O₂ allowing reduction and oxidation of the Pd sites readily. However, in the case of Pd/SOMS, the accessibility of Pd sites was significantly limited for O₂ species thereby showing lesser oxidation of Pd particles under the same reaction conditions compared to Pd/H-SOMS. For gas-phase HDC of TCE, higher TCE conversions were achieved over the Pd/H-SOMS catalyst compared to Pd/SOMS. In the presence of water, both Pd/SOMS and Pd/H-SOMS catalysts retained their catalytic activities, which was attributed to the hydrophobicity of the supports.

Acknowledgements

The financial support for this work was provided by the National Science Foundation through the Grant CBET-1436729. The authors also acknowledge partial support provided by the Ohio Coal Research Consortium.

MRCAT operations are supported by the Department of Energy and the MRCAT member institutions. This research used resources of the Advanced Photon Source, a U.S. Department of Energy (DOE) Office of Science User Facility operated for the DOE Office of Science by Argonne National Laboratory under Contract No. DE-AC02-06CH11357

References

- [1] B. Bakke, P.A. Stewart, M.A. Waters, Uses of and exposure to trichloroethylene in U.S. industry: a systematic literature review, *J. Occup. Environ. Hyg.* 4 (2007) 375–390.
- [2] R.E. Doherty, A history of the production and use of carbon tetrachloride, tetrachloroethylene, trichloroethylene and 1,1,1-trichloroethane in the United States: part 1—historical background; carbon tetrachloride and tetrachloroethylene, *Environ. Forensics* 1 (2000) 69–81.
- [3] U.S., E.P.A., Toxicological Review of Trichloroethylene in Support of Summary Information on the Integrated Risk Information System (Iris), U.S. Environmental Protection Agency, Washington, DC, 2011.
- [4] J. Hansen, et al., Risk of cancer among workers exposed to trichloroethylene: analysis of three nordic cohort studies, *J. Natl. Cancer Inst.* 105 (2013) 869–877.
- [5] M.J. Moran, J.S. Zogorski, P.J. Squillace, Chlorinated solvents in groundwater of the United States, *Environ. Sci. Technol.* 41 (2006) 74–81.
- [6] U.S., E.P.A., Occurrence Estimation Methodology and Occurrence Findings Report of the Six-Year Review of Existing National Primary Drinking Water Regulations; EPA-815/R-03-006; Washington, DC, 2003.
- [7] (ATSDR) A. f. T. S. a. D. R., in: PHS (Ed.), Toxicological Profile for Trichloroethylene (Tce) (Draft for Public Comment), U.S. Department of Health and Human Services, Atlanta, GA, 2014.
- [8] Y. Miyake, A. Sakoda, H. Yamanashi, H. Kaneda, M. Suzuki, Activated carbon adsorption of trichloroethylene (Tce) vapor stripped from tce-contaminated water, *Water Res.* 37 (2003) 1852–1858.
- [9] W.J. Blanford, E.J. Klingel, G.R. Johnson, R.B. Cain, C. Enfield, M.L. Brusseau, Performance Assessment of in-Well Aeration for the Remediation of an Aquifer Contaminated by a Multicomponent Immiscible liquid, *Soil Water Environ. Sci.* 725 (1999) 167–181.
- [10] H. Kim, D. Ahn, M.D. Annable, Enhanced removal of vocs from aquifers during air sparging using thickeners and surfactants: bench-scale experiments, *J. Contam. Hydrol.* 184 (2016) 25–34.
- [11] H.H. Russell, J.E. Matthews, G.W. Sewell, Tce Removal from Contaminated Soil and Ground Water Ground Water Issue, *Environ. Protect. Agency*, 1992, 12 pp.
- [12] P. Van der Avert, B.M. Weckhuysen, Low-temperature destruction of chlorinated hydrocarbons over lanthanide oxide based catalysts, *Angew. Chem. Int. Ed.* 41 (2002) 4730–4732.
- [13] Y. Nakano, L.Q. Hua, W. Nishijima, E. Shoto, M. Okada, Biodegradation of trichloroethylene (Tce) adsorbed on granular activated carbon (Gac), *Water Res.* 34 (2000) 4139–4142.
- [14] A.I. Zouboulis, P.A. Moussas, Groundwater and soil pollution: bioremediation, in: J.O. Nriagu (Ed.), *Encyclopedia of Environmental Health*, Elsevier, Burlington, 2011, pp. 1037–1044.
- [15] C. Liang, C.J. Bruell, M.C. Marley, K.L. Sperry, Persulfate oxidation for in situ remediation of tce: I. Activated by ferrous ion with and without a persulfate-thiosulfate redox couple, *Chemosphere* 55 (2004) 1213–1223.
- [16] X.D. Li, F.W. Schwartz, Dnapi remediation with in situ chemical oxidation using potassium permanganate, *J. Contam. Hydrol.* 68 (2004) 39–53.
- [17] S. Ordóñez, H. Sastre, F.V. Díez, Hydrodechlorination of aliphatic organochlorinated compounds over commercial hydrogenation catalysts, *Appl. Catal. B: Environ.* 25 (2000) 49–58.
- [18] M. Cobo, J. Becerra, M. Castelblanco, B. Cifuentes, J.A. Conesa, Catalytic hydrodechlorination of trichloroethylene in a novel naoh/2-propanol/methanol/water system on ceria-supported Pd and Rh catalysts, *J. Environ. Manage.* 158 (2015) 1–10.
- [19] N. Barrabeis, K. Föttinger, J. Llorca, A. Dafinov, F. Medina, J. Saí, C. Hardacre, G. n. Rupprechter, Pretreatment effect on Pt/CeO₂ catalyst in the selective hydrodechlorination of trichloroethylene, *J. Phys. Chem. C* 114 (2010) 17675–17682.
- [20] G.V. Lowry, M. Reinhard, Pd – catalyzed tce dechlorination in water: effect of [H₂](Aq) H₂ – utilizing competitive solutes on the tce dechlorination rate and product distribution, *Environ. Sci. Technol.* 35 (2001) 696–702.
- [21] M.G. Davie, H. Cheng, G.D. Hopkins, C.A. LeBron, M. Reinhard, Implementing heterogeneous catalytic dechlorination technology for remediating Tce-contaminated groundwater, *Environ. Sci. Technol.* 42 (2008) 8908–8915.
- [22] Juncheng Liu, F. He, Ed. Durham, D. Zhao, C.B. Roberts, Polysugar-stabilized Pd nanoparticles exhibiting high catalytic activities for hydrodechlorination of environmentally deleterious trichloroethylene, *Langmuir* 24 (2007) 328–336.
- [23] W. Sriwatanapongse, M. Reinhard, C.A. Klug, Reductive hydrodechlorination of trichloroethylene by palladium-on-alumina catalyst: 13c solid-state nmr study of surface reaction precursors, *Langmuir* 22 (2006) 4158–4164.
- [24] M.O. Nutt, K.N. Heck, P. Alvarez, M.S. Wong, Improved Pd-on-Au bimetallic nanoparticle catalysts for aqueous-phase trichloroethene hydrodechlorination, *Appl. Catal. B: Environ.* 69 (2006) 115–125.
- [25] E. Díaz, A. McCall, L. Faba, H. Sastre, S. Ordóñez, Trichloroethylene hydrodechlorination in water using formic acid as hydrogen. source: selection of catalyst and operation conditions, *Environ. Prog. Sustain. Energy* 32 (2013) 1217–1222.
- [26] N. Munakata, M. Reinhard, Palladium catalysis for the treatment of contaminated waters: a review, in: *Physicochemical Groundwater Remediation*, Springer, 2002, pp. 45–71.
- [27] S. Ordóñez, B.P. Vivas, F.V. Díez, Minimization of the deactivation of palladium catalysts in the hydrodechlorination of trichloroethylene in wastewaters, *Appl. Catal. B: Environ.* 95 (2010) 288–296.
- [28] X. Yu, T. Wu, X.J. Yang, J. Xu, J. Auzam, R. Semiat, Y.F. Han, Degradation of trichloroethylene by hydrodechlorination using formic acid as hydrogen source over supported Pd catalysts, *J. Hazard. Mater.* 305 (2016) 178–189.
- [29] A. Śrębowata, K. Tarach, V. Girmán, K. Góra-Marek, Catalytic removal of trichloroethylene from water over palladium loaded microporous and hierarchical zeolites, *Appl. Catal. B: Environ.* 181 (2016) 550–560.
- [30] G.V. Lowry, M. Reinhard, Hydrodehalogenation of 1-to 3-carbon halogenated organic compounds in water using a palladium catalyst and hydrogen gas, *Environ. Sci. Technol.* 33 (1999) 1905–1910.
- [31] M.O. Nutt, J.B. Hughes, M.S. Wong, Designing Pd-on-Au bimetallic nanoparticle catalysts for trichloroethene hydrodechlorination, *Environ. Sci. Technol.* 39 (2005) 1346–1353.
- [32] M. Zhang, D.B. Bacik, C.B. Roberts, D. Zhao, Catalytic hydrodechlorination of trichloroethylene in water with supported cmc-stabilized palladium nanoparticles, *Water Res.* 47 (2013) 3706–3715.
- [33] F. He, J. Liu, C.B. Roberts, D. Zhao, One-step 'Green' synthesis of Pd nanoparticles of controlled size and their catalytic activity for trichloroethene hydrodechlorination, *Ind. Eng. Chem. Res.* 48 (2009) 6550–6557.
- [34] R. Navon, S. Eldad, K. Mackenzie, F.-D. Kopinke, Protection of palladium catalysts for hydrodechlorination of chlorinated organic compounds in wastewaters, *Appl. Catal. B: Environ.* 119–120 (2012) 241–247.
- [35] F.-D. Kopinke, D. Angeles-Wedler, D. Fritsch, K. Mackenzie, Pd-catalyzed hydrodechlorination of chlorinated aromatics in contaminated Waters—Effects of surfactants, organic matter and catalyst protection by silicone coating, *Appl. Catal. B: Environ.* 96 (2010) 323–328.
- [36] D. Comandella, S. Wosidlo, A. Georgi, F.-D. Kopinke, K. Mackenzie, Efforts for long-term protection of palladium hydrodechlorination catalysts, *Appl. Catal. B: Environ.* 186 (2016) 204–211.
- [37] J. Wei, Y. Qian, L. Wang, Y. Ge, L. Su, D. Zhai, J. Wang, J. Yu, Enhancement of degradation and dechlorination of trichloroethylene via supporting palladium/iron bimetallic nanoparticles onto mesoporous silica, *Catalysts* 6 (2016) 105.
- [38] Y. Han, C. Liu, J. Horita, W. Yan, Trichloroethene hydrodechlorination by Pd-Fe bimetallic nanoparticles: solute-induced catalyst deactivation analyzed by carbon isotope fractionation, *Appl. Catal. B: Environ.* 188 (2016) 77–86.
- [39] X.-q. Li, D.W. Elliott, W.x. Zhang, Zero-valent iron nanoparticles for abatement of environmental pollutants: materials and engineering aspects, *Crit. Rev. Solid State Mater. Sci.* 31 (2006) 111–122.
- [40] H.-L. Lien, W.-X. Zhang, Nanoscale Pd/Fe bimetallic particles: catalytic effects of palladium on hydrodechlorination, *Appl. Catal. B: Environ.* 77 (2007) 110–116.

- [41] L. Wu, S.M. Ritchie, Removal of trichloroethylene from water by cellulose acetate supported bimetallic Ni/Fe nanoparticles, *Chemosphere* 63 (2006) 285–292.
- [42] C.J. Lin, S.L. Lo, Y.H. Liou, Dechlorination of trichloroethylene in aqueous solution by noble metal-modified iron, *J. Hazard. Mater.* 116 (2004) 219–228.
- [43] Y.-H. Tee, E. Grulke, D. Bhattacharyya, Role of Ni/Fe nanoparticle composition on the degradation of trichloroethylene from water, *Ind. Eng. Chem. Res.* 44 (2005) 7062–7070.
- [44] Q. Guo, L. Ren, Hydrodechlorination of trichloroethylene over mop/ Γ -Al₂O₃ catalyst with high surface area, *Catal. Today* 264 (2016) 158–162.
- [45] X. Liu, L. Ren, The influence of polyethylene glycol on the synthesis and activity of mop for the hydrodechlorination of trichloroethylene, *RSC Adv.* 6 (2016) 7413–7418.
- [46] E. Finocchio, G. Sapienza, M. Baldi, G. Busca, Trichloroethylene catalytic conversion over acidic solid catalysts, *Appl. Catal. B: Environ.* 51 (2004) 143–148.
- [47] B.T. Meshesha, N. Barrabés, J. Llorca, A. Dafinov, F. Medina, K. Föttinger, PdCu alloy nanoparticles on alumina as selective catalysts for trichloroethylene hydrodechlorination to ethylene, *Appl. Catal. A: Gen.* 453 (2013) 130–141.
- [48] B.T. Meshesha, N. Barrabés, K. Föttinger, R.J. Chimentão, J. Llorca, F. Medina, G. Rupprechter, J.E. Sueiras, Gas-phase hydrodechlorination of trichloroethylene over Pd/Nimgal mixed oxide catalysts, *Appl. Catal. B: Environ.* 117–118 (2012) 236–245.
- [49] H. Sohn, G. Celik, S. Gunduz, S.L. Dean, E. Painting, P.L. Edmiston, U.S. Ozkan, Hydrodechlorination of trichloroethylene over Pd supported on swellable organically-modified silica (Soms), *Appl. Catal. B: Environ.* 203 (2017) 641–653.
- [50] P.L. Edmiston, L.A. Underwood, Absorption of dissolved organic species from water using organically modified silica that swells, *Sep. Purif. Technol.* 66 (2009) 532–540.
- [51] C.M. Burkett, P.L. Edmiston, Highly swellable sol–gels prepared by chemical modification of silanol groups prior to drying, *J. Non-Cryst. Solids* 351 (2005) 3174–3178.
- [52] C.M. Burkett, L.A. Underwood, R.S. Volzer, J.A. Baughman, P.L. Edmiston, Organic–inorganic hybrid materials that rapidly swell in non-polar liquids: nanoscale morphology and swelling mechanism, *Chem. Mater.* 20 (2008) 1312–1321.
- [53] P.L. Edmiston, Swellable Sol–Gels, *Methods of Making, and Use Thereof* US8119759 B2, 2012.
- [54] P.L. Edmiston, Swellable Materials and Methods of Use US 8 367, 793 B2, 2013.
- [55] P.J. Launer, Infrared analysis of organosilicon compounds: spectra-structure correlations, *Silicone Compd. Regist. Rev.* (1987) 100.
- [56] R.M. Almeida, C.G. Pantano, Structural investigation of silica gel films by infrared spectroscopy, *J. Appl. Phys.* 68 (1990) 4225–4232.
- [57] F. Galeener, G. Lucovsky, Longitudinal optical vibrations in glasses: GeO_2 and SiO_2 , *Phys. Rev. Lett.* 37 (1976) 1474–1478.
- [58] M. Payne, J. Inkson, Longitudinal-optic-transverse-optic vibrational mode splittings in tetrahedral network glasses, *J. Non-Cryst. Solids* 68 (1984) 351–360.
- [59] S. De Leeuw, M.F. Thorpe, Coulomb splittings in glasses, *Phys. Rev. Lett.* 55 (1985) 2879.
- [60] D. Lin-Vien, N.B. Colthup, W.G. Fateley, J.G. Grasselli, *The Handbook of Infrared and Raman Characteristic Frequencies of Organic Molecules*, Elsevier, 1991, 2017.
- [61] F. Iacopi, G. Beyer, Y. Travaly, C. Waldfried, D.M. Gage, R.H. Dauskardt, K. Houthoofd, P. Jacobs, P. Adriaenssens, K. Schulze, Thermomechanical properties of thin organosilicate glass films treated with ultraviolet-assisted cure, *Acta Mater.* 55 (2007) 1407–1414.
- [62] R. Rulkens, Y. Ni, I. Manners, Living anionic ring-opening polymerization of silicon-bridged 1 ferrocenophanes: synthesis and characterization of poly(ferrocenylsilane)-polysiloxane block copolymers, *J. Am. Chem. Soc.* 116 (1994) 12121–12122.
- [63] P.-Y. Mabboux, K.K. Gleason, Chemical bonding structure of low dielectric constant Si:O:C:H films characterized by solid-state Nmr, *J. Electrochem. Soc.* 152 (2005) F7–F13.
- [64] J.J. Yang, I.M. El-Nahhal, I.-S. Chuang, G.E. Maciel, Synthesis and solid-State nmr structural characterization of polysiloxane-immobilized amine ligands and their metal complexes, *J. Non-Cryst. Solids* 209 (1997) 19–39.
- [65] A. Shimojima, D. Mochizuki, K. Kuroda, Synthesis of silylated derivatives of a layered polysilicate kanemite with mono-, di-, and trichloro (Alkyl) silanes, *Chem. Mater.* 13 (2001) 3603–3609.
- [66] S. Choi, Y. Kim, I. Kim, C.-S. Ha, Effect of organosilica isomers on the interfacial interaction in polyimide/Aromatic organosilica hybrids, *J. Appl. Polym. Sci.* 103 (2007) 2507–2513.
- [67] P. Kast, M. Friedrich, D. Teschner, F. Girgsdies, T. Lunkenbein, R. Naumann d'Alnoncourt, M. Behrens, R. Schlögl, Co oxidation as a test reaction for strong metal–support interaction in nanostructured Pd/Feox powder catalysts, *Appl. Catal. A: Gen.* 502 (2015) 8–17.
- [68] N. Mahata, V. Vishwanathan, Influence of palladium precursors on structural properties and phenol hydrogenation characteristics of supported palladium catalysts, *J. Catal.* 196 (2000) 262–270.
- [69] S. Somboonthanakij, O. Mekasuwandumrong, J. Panpranot, T. Nimmanwudtipong, R. Strobel, S.E. Pratsinis, P. Praserthdam, Characteristics and catalytic properties of Pd/SiO₂ synthesized by one-step flame spray pyrolysis in liquid-phase hydrogenation of 1-heptyne, *Catal. Lett.* 119 (2007) 346–352.
- [70] O. Verho, H. Zheng, K.P.J. Gustafson, A. Nagendiran, X. Zou, J.-E. Bäckvall, Application of Pd nanoparticles supported on mesoporous hollow silica nanospheres for the efficient and selective semihydrogenation of alkynes, *ChemCatChem* 8 (2016) 773–778.
- [71] K. Wu, X. Qian, L. Chen, Z. Xu, S. Zheng, D. Zhu, Effective liquid phase hydrodechlorination of diclofenac catalysed by Pd/CeO₂, *RSC Adv.* 5 (2015) 18702–18709.
- [72] B. Schrick, J.L. Blough, A.D. Jones, T.E. Mallouk, Hydrodechlorination of trichloroethylene to hydrocarbons using bimetallic nickel-iron nanoparticles, *Chem. Mater.* 14 (2002) 5140–5147.